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2 **Effects of wheat stubble on runoff, infiltration, and erosion of farmland in the Loess**
3 **Plateau, China subjected to simulated rainfall**

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11 **Abstract**

12 Soil and water losses in agriculture are major environmental problems worldwide, especially on
13 the Loess Plateau, China. Summer fallow management may help to control soil erosion and conserve
14 water. This study investigated the effects of wheat stubble on runoff, infiltration, and soil loss in
15 laboratory plots under simulated rainfall. The treatments comprised wheat stubble cover (WS) and
16 traditional plowing (TP) in runoff plots (4.0 m × 1.0 m) with three slope gradients (5°, 10°, and 15°)
17 under simulated rainfall at 80 mm h⁻¹ for 1 h. The runoff volume from WS plots was significantly less
18 than that from TP. The runoff reduction with WS ranged from 91.92–92.83% compared with TP. The



19 runoff rates varied with the runoff volume in the same manner. Under WS, sediment losses (2.41–3.78 g
20 m^{-2}) were reduced dramatically compared with TP (304.31–731.23 g m^{-2}). The sediment concentration
21 was also significantly lower with WS than TP. The infiltration amount was higher with WS (94.8–96.2%
22 of rainwater infiltrated) than TP (35.4–57.1%). Thus, stubble cover can help to control erosion and
23 conserve soil and water resources.

24 **Keywords:** Loess Plateau, Soil erosion, Summer fallow, Traditional plowing, Wheat stubble,
25

26 1. Introduction

27 Soil and water losses from agricultural land, particularly sloping farmland, are regarded as major
28 environmental threats to ecosystem sustainability on the Loess Plateau, China. Approximately 60% of
29 the total watershed sediment and runoff is derived from sloped farmland due to natural and human
30 factors, such as the precipitation intensity, geomorphology, and soil management practices, which all
31 contribute to farmland degradation (Fu et al., 2000; Fu et al., 2006; Kang et al., 2001; Shi and Shao,
32 2000).

33 The susceptibility to soil and water losses is higher on farmland than land with other vegetation
34 types (e.g., forestry, shrub, and grass) (Boardman et al., 1990; Shi and Shao, 2000; Tang, 2004). This is
35 because the characteristics of crops, including the crop canopy architecture and root system, differ from
36 those of other vegetation types, which leads to severe soil erosion on farmland (Cerdà et al., 2009;
37 Gómez et al., 2004; Llorens and Domingo, 2007). In particular, crops are harvested, which means that
38 there is a fallow period when vegetation does not cover the soil surface. In the Loess Plateau region of



39 China, farmers conduct tillage practices during the summer fallow period after harvesting winter wheat
40 in June. The main aims of summer tillage are removing weeds, and creating a favorable rough surface to
41 maximize rainfall capture and minimize soil evaporation (Hammel et al., 1981; Vermang et al., 2015).
42 However, soil management in the summer fallow period may completely disrupt the surface soil,
43 thereby making farmland more susceptible to severe soil erosion in the fallow rainy season (Wang et al.,
44 2016). In the Loess Plateau region, most of the annual rainfall is concentrated in the summer between
45 July and September, when 60–70% of the total annual rainfall occurs (Shi and Shao, 2000). Thus, the
46 rainfall is erratic and it occurs with a high intensity and short duration. Summer tillage of bare sloped
47 farmland means that extreme erosive rainfall events can cause severe surface runoff and soil erosion.
48 Thus, the occurrence of rainfall and the tillage of sloped farmland in the summer contribute greatly to
49 extreme soil erosion. This is why soil erosion from sloped farmland is recognized as the main source of
50 sediment losses, which are higher than those from other land use types. Therefore, it is crucial to prevent
51 soil erosion from sloped farmland during the summer fallow period.

52 To reduce soil erosion from farmland, numerous studies have considered the roles of surface
53 cover and soil management practices in conservation agriculture worldwide (Bescansa et al., 2006;
54 Jordán et al., 2010; Prosdocimi et al., 2016; Swella et al., 2015; Won et al., 2012). Retaining a surface
55 covered with a layer of crop residues is a suitable management practice for preventing soil losses and
56 conserving rainwater on farmland. Thus, Gholami et al. (2013) investigated the effects of straw mulch
57 on soil erosion under various rainfall intensities in laboratory simulations, where the results showed that
58 straw mulch was effective in delaying the runoff initiation time, as well as reducing splash erosion, the
59 runoff volume, and sediment losses. Kukal and Sarkar (2010) also investigated the effects of wheat



60 mulch on splash erosion and infiltration under simulated rainfall conditions, and concluded that straw
61 mulch can play an important role in mitigating splash erosion and increasing infiltration. Jordán et al.
62 (2010) showed that the application of wheat straw mulch could improve the physical and chemical
63 properties of soil, as well as reduce runoff and soil losses in cultivated land under semiarid conditions.
64 Prosdocimi et al. (2016) examined the effects of straw mulch on soil erodibility and surface runoff, and
65 found that straw cover can be effective in reducing soil erodibility, thereby decreasing the soil erosion
66 risk on cultivated land. Similar results were obtained by Cerdà et al. (2016), who found that straw
67 residues were effective in reducing soil and water losses from agricultural land under simulated rainfall
68 in the field. Swella et al. (2015) showed that retaining crop residues increased rainfall infiltration and
69 reduced evaporation during the summer and autumn. Aboudrare et al. (2006) investigated the effects of
70 fallow management on soil water storage during the fallow period in semiarid Mediterranean conditions,
71 where they found that the implementation of appropriate management could maximize the capture of
72 rainwater. Blanco and Lal (2008) investigated tillage practices to increase rainwater storage for winter
73 wheat production and runoff reduction in the dryland areas of the western USA and Great Plains during
74 the summer fallow periods. Lee and Yang (1965) reported that summer plowing during the summer
75 fallow period can improve the physical, chemical, and biological properties of soil, thereby increasing
76 winter wheat production in the Loess Plateau region. These studies have provided insights into the
77 importance of surface cover and soil management practice for controlling soil and water losses from
78 farmland. Beneficial strategies for soil and water conservation are based on the following principles:
79 providing cover to absorb raindrop kinetic energy as well as reducing splash erosion and the overland
80 flow velocity (Gholami et al., 2014; Gholami et al., 2013; Sadeghi et al., 2015); improving the physical



81 and chemical properties of soil, such as the organic matter content and stability of aggregates (Jordán et
82 al., 2010); increasing the soil infiltration capacity and decreasing evaporation (Adekalu et al., 2007;
83 Swella et al., 2015; Todd et al., 1991); and creating a rough surface to decrease runoff and enhance
84 infiltration (Strudley et al., 2008; Vermang et al., 2015; Wang et al., 2016). The obvious advantages of
85 crop residues as a soil cover in conservation agriculture are known, but the effects of crop residues on
86 rainwater capture and soil and water losses from sloped farmland have been studied little, especially
87 during the summer fallow period in the Loess Plateau region.

88 On the Loess Plateau, the area of cultivated land is 145 800 km², which accounts for 25.6% of
89 the total land area. About 70% of the cultivated land comprises rainfed areas, which are distributed in
90 mountainous and hilly regions (NDRC et al., 2010). On sloped farmland, a combination of continuous
91 and intense cultivation, inappropriate soil management, and concentrated rainfall in the summer fallow
92 period have caused severe soil erosion, thereby decreasing the soil productivity and increasing land
93 degradation. Wheat is one of the major crops in the Loess Plateau region, where it accounts about 35%
94 of the total cultivated area and 30% of the total crop production (NDRC et al., 2010). The availability
95 and costs of application with stubble cover mean that it is practical for farmers to implement this method
96 in the summer fallow periods. In addition, retaining stubble cover is an efficient and environmentally
97 friendly method that utilizes the crop biomass. The present study aimed to investigate the effects of
98 retaining the stubble from wheat on runoff, erosion, and rainwater capture in a laboratory plot under
99 simulated rainfall and different slope conditions. The results of this study provide a better understanding
100 of the effects of stubble cover for farmers and policy makers, which may be important for conserving
101 soil productivity and supporting sustainable agriculture in the Loess Plateau rainfed area.



102 **2. Material and Methods**

103 2.1 Experimental plots and rainfall simulator

104 This study was performed at the Laboratory of Soil and Water Conservation located at the
105 Northwest A&F University campus, Yangling, Shannxi Province. Yangling (E107°59'–108°08',
106 N34°14'–34°20') is located on the Guanzhong Plain at an altitude of 516.4–540.1 m. This area has a
107 warm temperate, semi-humid monsoon climate with an average annual temperature of 12.9°C. The
108 annual precipitation is 635.1 mm and the highest rainfall occurs in the period from July to September,
109 which accounts for 60–70% of the annual precipitation. The experiments were conducted on runoff plots
110 built in 2009, as shown in Fig. 1a. The areas of most plots used for laboratory rainfall simulations in
111 previous studies were less than 5 m² (Huang et al., 2013; Wu et al., 2014; Zhao et al., 2013). In the
112 present study, the runoff plot measured 4.0 m (length) × 1.0 m (width) × 0.6 m (depth), and four runoff
113 plots comprised a slope gradient group. The slope gradients of the runoff plots were 5°, 10°, and 15°,
114 which represented farmland with slight, gentle, and steep slopes in the field based on the classification
115 of farmland in the Loess Plateau region, where 42% of the farmland has slopes of 5–15° (NDRC et al.,
116 2010). The runoff plots were filled with soil, which was taken from the top 0–20 cm soil layer of a farm
117 in Yangling after the runoff plot was constructed in July, 2009. The soil was clay loam and its major
118 physico-chemical properties are summarized in Table 1. Before placing the soil in the plot, a 10-cm
119 layer of sand was laid at the bottom to allow free drainage. After filling with soil, the runoff plots were
120 left for one year to allow natural compaction in order to obtain soil properties similar to the natural



121 conditions. Each runoff plot had an aluminum sheet at the lower end of plot, which served as an outlet
122 for collecting runoff samples.

123 Rainfall simulation is used widely as a method for studying runoff and erosion processes. The
124 portable rainfall simulation system used in the present study was developed by the Institute of Soil and
125 Water Conservation, Chinese Academy of Science and Ministry of Water Resources, as described by
126 Wang et al. (2016). Its main components comprised a pumping system, inlet pipes, control valve, steel
127 pipes, piezometer, spray nozzles and bracket to hold the spray nozzle (Fig. 1b). Raindrops were
128 generated by a spray nozzle with a drop height of 7.5 m and they had comparable characteristics to
129 natural rainfall in terms of height. Depending on the water supplied by the pump, the rainfall intensity
130 varied according to different pressures displayed on the piezometer, which was installed at the inlet of
131 the steel pipe. The rainfall simulator system was calibrated by the pressure to obtain different intensities.
132 The effective cover area of the simulated rainfall was 5.0 m in length and 4.0 m in width, which was
133 sufficient to cover the area of the two runoff plots while avoiding border interference. Therefore,
134 simulated rainfall was applied simultaneously to two neighboring plots. The two rainfall simulators were
135 placed between neighboring runoff plots. One rainfall simulator was placed 0.5 m from the upslope plot
136 edge and the other was placed 0.5 m from the downslope edge. During rainfall simulations, the adjacent
137 plots were covered with plastic sheets to prevent rainfall falling on the soil.

138 2.2 Experimental treatments

139 The experiment was conducted in June 2013. The wheat variety used was Xiaoyan-22 and it was
140 sown in four plots during October 2012. Soybean (2010) and maize (2011) had been planted in the plots



141 for scientific research before the plots were planted with wheat. The plots were cleaned, plowed, and
142 prepared to obtain a seed bed, as shown in Fig. 1a. The wheat was sown in rows with a space of 20 cm
143 and the sowing rate was 13 g m^{-2} . During the wheat growth season, the crop management practices were
144 similar to those employed by local farmers in the field. The mature wheat was harvested in June 2013.
145 The two plot treatments comprised wheat stubble (WS) with a height 20 cm above the ground, whereas
146 the aboveground parts of the wheat biomass were cleared from other two plots and traditional plowing
147 was applied (TP). TP is typically used by local farmers on the sloping farmland in the Loess Plateau
148 region. TP uses a single plow with an iron frame and an attached blade to break up and turn the soil
149 across the slope direction at a depth of 20 cm.

150 2.3 Rainfall simulation and data analysis

151 After preparing all the treatments, the plot border was hydrologically isolated with plastic boards
152 inserted 15 cm underground and 15 cm aboveground, which prevented runoff flowing out or into
153 adjacent plots. A 60-min rainfall simulation at an intensity of 80 mm h^{-1} was used in this experiment
154 based on long-term monitoring result of the natural rainfall intensity on the Loess Plateau. After starting
155 a rainfall simulation, the time to runoff initiation was recorded, which was determined when runoff
156 started to flow from the outlet of the plot. Runoff samples were collected at intervals of 2.0 min from
157 two plots with plastic buckets, which had been weighed previously. After finishing a rainfall simulation,
158 the samples were weighed and left undisturbed for 24 h. The deposited sediment was then poured into
159 aluminum boxes, which had been weighed previously, and the sediment was oven dried at 105°C for 24
160 h and then weighed again. Based on the runoff volume, sediment yield, and time interval data, the runoff



161 rate (mm min^{-1}), infiltration rate (mm min^{-1}), cumulative infiltration amount (mm), sediment
162 concentration (g L^{-1}), and sediment loss (g m^{-2}) were calculated (Zhao et al., 2014).

163 One-way analysis of variance was used to analyze the effects of different treatments on the
164 runoff rate, runoff volume, sediment concentration and loss, and the infiltration amount. Statistical
165 analyses were performed using IBM SPSS Statistics 19.0 (IBM, 2010). The figures were drawn using
166 Sigma Plot 10.0 (Systat, 2008).

167 3. Results and discussion

168 3.1 Runoff

169 Table 1 summarizes the time to runoff initiation, runoff rate, and runoff volume for different
170 treatment plots. Figure 2 shows the dynamics of the runoff rate during the rainfall simulations. WS
171 delayed the runoff initiation time by about 4–18 min compared with TP, which indicates that WS had a
172 positive effect on runoff generation. In general, the runoff initiation time decrease in the order of WS >
173 TP with the three slope gradients, and there was a significant difference between the two treatments ($P <$
174 0.05). In addition, the runoff initiation time decreased as the slope increased, as shown by Yair and
175 Lavee (1976). However, the difference in the runoff initiation time between WS and TP decreased as the
176 slope increased, although the difference was not significant. For example, the difference was 18.9 min at
177 5° and 4.06 min in 15° . Thus, WS cover had limited effects on delaying the initiation of runoff with
178 relatively steep slopes.

179 WS had a lower runoff rate and runoff volume, as shown in Table 1. There were significant
180 differences between WS and TP in terms of both the runoff rate and runoff volume ($P < 0.05$). The



181 runoff volumes were 37.86, 47.28, and 51.42 mm under TP with the three slopes of 5°, 10°, and 15°,
182 respectively, but only 3.00, 3.39, and 4.16 mm under WS. The runoff volume under WS was reduced by
183 91.92–92.83% compared with that under TP. The runoff rate varied in the same manner as the runoff
184 volume. The dynamics of the runoff rate were also captured, and the runoff processes varied greatly
185 among treatments. In general, the runoff rate response seemed to be less sensitive to rainfall, where there
186 was a slight increasing trend in the initial runoff stages, before remaining at low values during the
187 rainfall simulation under WS compared with TP. Under TP, the runoff rate increased during the first 25–
188 30 min, before stabilizing after 40 min with all three slopes. These results demonstrate that WS direct
189 affected the delay in runoff generation and reduced the amount of runoff. These findings are similar to
190 the results reported by Jordán et al. (2010), who found that mulch cover significantly reduced the runoff
191 compared with bare slope. Similarly, Puustinen et al. (2005) found that the presence of mulch facilitated
192 infiltration and delayed runoff, thereby resulting in much less runoff. In addition, Won et al. (2012)
193 found that straw mat cover resulted in significant less runoff because the residues acted as a barrier to
194 reduce the overland flow velocity, thereby allowing more rainfall to infiltrate, which minimized the
195 runoff rate and reduced the runoff volume. Therefore, these results demonstrate that WS can
196 significantly reduce runoff.

197 3.2 Soil infiltration

198 The soil water content is a limiting factor that affects crop yields in rainfed farmland in semiarid
199 regions, especially on the Loess Plateau. Tillage during summer fallow in the Loess Plateau region aims
200 to retain rainwater and conserve soil water moisture for the next season's crop. Precipitation is the main
201 source of soil moisture for agronomic crops in this area and it measures the proportion of rainwater that



202 infiltrates into the soil, thereby providing insights into the effects of WS cover on soil water
203 conservation.

204 Figure 3 shows the dynamic infiltration processes for all the treatments with various slope
205 gradients under simulated rainfall, which indicates that there was a higher initial infiltration rate under
206 TP, but it decreased rapidly as the simulated rainfall proceeded and reached a steady infiltration rate with
207 small fluctuations at 25–30 min, as observed by Wang et al. (2016). This may be explained by plowing
208 producing a loose and rough soil surface. In the initial rainfall period, the soil particles were splashed
209 and rainfall water was stored in micro-depressions. Therefore, the initial infiltration rate was higher and
210 the runoff rate was lower, as shown by the results in Fig. 2 and Fig. 3. As the rainfall proceeded, the
211 surface was sealed by the impact of raindrops, so the infiltration rate decreased rapidly (Bissonnais,
212 1996; Shen et al., 2016). However, the infiltration rate remained at a higher level under WS compared
213 with TP in all of the rainfall simulation events under three slope gradients.

214 The cumulative infiltration amount showed that WS cover was an effective method for capturing
215 rainfall water. Table 1 shows that the cumulative infiltration amount was significantly higher under WS
216 than TP. The cumulative infiltration amounts under WS were 76.11, 75.95, and 75.83 mm with slopes of
217 5°, 10°, and 15°, respectively, i.e., 96.2%, 95.7%, and 94.8% rainfall infiltration, and thus there were
218 slight decreases as the slope gradient increased. A similar trend was also observed under TP, where the
219 cumulative infiltration amounts were 43.90, 32.53, and 28.14 mm with slopes of 5°, 10°, and 15°,
220 respectively, i.e., 53.7%, 40.8%, and 35.4% rainfall infiltration. According to these results, the
221 percentage of rainwater infiltrated indicated that WS cover significantly improved the capture of



222 rainwater. The standing stubble at the soil-atmosphere interface protected the soil surface from the direct
223 impact of raindrops and intercepted the rainfall near the soil surface, thereby give the rainwater more
224 time for infiltration. This might also explain the longer time required to generate runoff, the lower runoff
225 rates, and the higher infiltration amount under WS. Similar findings were obtained by Prosdocimi et al.
226 (2016) using barley straw mulch in Mediterranean vineyards, Jordán et al. (2010) using wheat straw
227 mulch under semiarid condition in southern Spain, Won et al. (2012) with straw mats covering soil in
228 laboratory rainfall simulations, Moreno-Ramón et al. (2014) in soil mulched with coffee husks in
229 laboratory rainfall simulations, and by Swella et al. (2015) using standing residues in a rainfed
230 agriculture system.

231 3.3 Soil loss

232 Table 1 shows the sediment concentration and total sediment loss in different treatment plots.

233 The sediment concentration under TP ranged between 8.14–14.90 g L⁻¹, which was significantly higher
234 than that under WS, i.e., 0.82–1.01 g L⁻¹. The total sediment loss varied in the same manner. The total
235 sediment loss under WS was 2.41–3.78 g m⁻², which was much lower than that under TP, i.e., 304.31–
236 731.23 g m⁻², and the sediment loss increased with the slope gradient. These results are consistent with
237 those expected because the stubble cover decreased soil losses (Hueso-González et al., 2015). Figure 4
238 also shows that the dynamic changes in the sediment concentration decreased dramatically under WS
239 compared with TP. Under TP, the sediment concentration increased rapidly in the first few minutes
240 because runoff was generated from the loose particles on the soil surface, before decreasing slightly. As
241 the rainfall simulation progressed, the sediment concentration reached a stable state because the loose
242 particles were exhausted. The observed changes in the sediment concentration during a rainfall event



243 followed a typical pattern, as reported by Jordán et al. (2010), Roth and Helming (1992), and Shen et al.
244 (2016). The differences in behavior between WS and TP may have been due to the standing stubble,
245 which absorbed the kinetic energy of raindrops and prevented their direct splashing (Gholami et al.,
246 2013). In addition, Jordán et al. (2010) noted that soil erosion was substantially reduced by high straw
247 residues due to decreased runoff detachment and an increased infiltration rate. Moreover, Won et al.
248 (2012) reported that substantial decreases in the sediment concentration under treatment with straw mat
249 cover was due to reduced runoff, thereby minimizing the overland flow transport capacity. Furthermore,
250 the wheat roots played an important role by reinforcing the soil and increasing infiltration (Katuwal et
251 al., 2013; Shinohara et al., 2016), which reduced the sediment concentration and sediment losses under
252 WS. Our results showed that TP disturbed the surface soil, which increased the available sediment
253 source and the likelihood of soil being detached and transported by raindrops and overland flow.
254 Ultimately, this may explain the higher sediment concentration under TP. These results agree with those
255 obtained by Engel et al. (2009), who found that tillage disturbed the soil surface, thereby facilitating the
256 transport of soil particles via overland flow. Celik (2005) also concluded that disturbance by tillage
257 could decrease the stability of soil particles so they were more likely to be detached. Vermang et al.
258 (2015) found that a rougher soil surface yielded a significantly higher sediment concentration because
259 the concentrated runoff could transport more particles.

260 3.4 Implications

261 The Loess Plateau is highly susceptible to soil erosion because of its erodible soil, sparse
262 vegetation cover, and concentrated summer precipitation. To address this problem, the Chinese
263 government launched the “Grain-for-Green” project in 1999 with the aim of converting steep sloping



264 (>25°) farmland into forest or grassland. However, 20% of the total farmland still has slopes of 15–25°
265 and 7% has slopes larger than 25° (NDRC et al., 2010). Our results confirm that WS delayed the runoff,
266 reduced soil erosion, and increased rainwater infiltration compared with TP. Figure 5 shows clearly that
267 WS was also effective in reducing the runoff rate and sediment concentration compared with TP. WS
268 significantly reduced the average runoff rate from 0.79 mm min⁻¹ (TP) to 0.08 mm min⁻¹ (WS), and the
269 average sediment concentration from 11.12 g L⁻¹ (TP) to 0.91 g L⁻¹ (WS). Therefore, the average
270 reductions in the runoff volume and sediment loss under WS were 92.27% and 99.39%, respectively,
271 compared with TP. Therefore, these results indicate that retaining stubble at the soil-atmosphere
272 interface can control soil erosion from farmland during the summer fallow period.

273 Furthermore, these experiments were conducted using small laboratory plots to test the
274 immediate effects of stubble on runoff and soil erosion under simulated rainfall conditions. It should be
275 noted that scale issues are important when interpreting results that affect practical applications (Sadeghi
276 et al., 2015). However, the results obtained in the present study can be used as comparative information
277 to indicate how WS may reduce the risk of soil erosion and increase rainwater capture on sloped
278 farmland. Mulumba and Lal (2008) noted that the WS cover varied according to the site-specific
279 conditions, so more studies are necessary to investigate a wider range of slopes, standing stubble height,
280 and different plot and field scales. Furthermore, it is necessary to determine the long-term impacts of
281 stubble cover on the soil properties (Bescansa et al., 2006; Lipiec et al., 2006), microbial activity (Song
282 et al., 2002), soil carbon content (Amundson et al., 2015; Lal, 1993; Van et al., 2007), and soil nutrient
283 cycling (Holland, 2004), as well as its relationships with soil erosion, crop productivity, and crop



284 biomass utilization. These studies may provide farmers and policymakers with a comprehensive
285 understanding of crop residue management under specific conditions in the Loess Plateau region.

286 **4. Conclusion**

287 This study investigated the effects of WS and TP on runoff and soil erosion in laboratory plots
288 under simulated rainfall with three different slope gradients. The results showed clearly that stubble can
289 be used as an effective management practice during the summer fallow period in the Loess Plateau
290 region. WS delayed the runoff initiation time, as well as decreased the runoff rate and runoff volume.
291 The sediment concentration and sediment losses were also decreased by WS compared with TP. In
292 addition, WS was highly beneficial because it absorbed the kinetic energy of raindrops and promoted
293 water infiltration. Thus, WS performed well according to all the variables considered in this study. The
294 reductions in runoff and the sediment concentration under WS were 91.0–93.2% and 99.2–99.6%,
295 respectively, compared with TP, and infiltration was 1.69–2.45 times higher under WS than TP. In
296 conclusion, WS was beneficial for reducing runoff and sediment losses, as well as increasing infiltration
297 under simulated rainfall. Stubble cover can be adopted as a management practice by farmers to control
298 soil erosion and promote rainwater conservation in the summer fallow period. Future studies should
299 evaluate the performance of this crop biomass by-product in diverse field conditions.

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304 **References**

- 305 Aboudrare, A., Debaeke, P., Bouaziz, A., Chekli, H., 2006. Effects of soil tillage and fallow
306 management on soil water storage and sunflower production in a semi-arid Mediterranean climate.
307 *Agricultural Water Management* 83, 183-196.
- 308 Adekalu, K.O., Olorunfemi, I.A., Osunbitan, J.A., 2007. Grass mulching effect on infiltration, surface
309 runoff and soil loss of three agricultural soils in Nigeria. *Bioresource technology* 98, 912-917.
- 310 Amundson, R., Berhe, A.A., Hopmans, J.W., Olson, C., Sztein, A.E., Sparks, D.L., 2015. Soil and
311 human security in the 21st century. *Science* 348 (6235),1261071-6.
- 312 Bescansa, P., Imaz, M.J., Virto, I., Enrique, A., Hoogmoed, W.B., 2006. Soil water retention as affected
313 by tillage and residue management in semiarid Spain. *Soil and Tillage Research* 87, 19-27.
- 314 Bissonnais, Y.L., 1996. Aggregate stability and assessment of soil crustability and erodibility: I. theory
315 and methodology. *European Journal of Soil Science* 47, 425-437.
- 316 Blanco, H., Lal, R., 2008. Principles of soil conservation and management. Springer p167-169.
- 317 Boardman, J., Foster, I.D.L., Dearing, J.A., 1990. *Soil Erosion on Agricultural Land*. John Wiley and
318 Sons. Chichester.
- 319 Celik, I., 2005. Land-use effects on organic matter and physical properties of soil in a southern
320 Mediterranean highland of Turkey. *Soil and Tillage Research* 83, 270-277.
- 321 Cerdà, A., González-Pelayo, Ó., Giménez-Morera, A., Jordán, A., Pereira, P., Novara, A., Brevik, E.C.,
322 Prosdocimi, M., Mahmoodabadi, M., Keesstra, S., Orenes, F.G., Ritsema, C.J., 2016. Use of barley
323 straw residues to avoid high erosion and runoff rates on persimmon plantations in Eastern Spain under
324 low frequency–high magnitude simulated rainfall events. *Soil Research* 54, 154-165.



- 325 Cerdà, A., Morera, A.G., Bodí, M.B., 2009. Soil and water losses from new citrus orchards growing on
326 sloped soils in the western Mediterranean basin. *Earth Surface Processes and Landforms* 34, 1822-1830.
- 327 Engel, F.L., Bertol, I., Ritter, S.R., Paz González, A., Paz-Ferreiro, J., Vidal Vázquez, E., 2009. Soil
328 erosion under simulated rainfall in relation to phenological stages of soybeans and tillage methods in
329 Lages, SC, Brazil. *Soil and Tillage Research* 103, 216-221.
- 330 Fu, B., Chen, L., Ma, K., Zhou, H., Wang, J., 2000. The relationships between land use and soil
331 conditions in the hilly area of the loess plateau in northern Shaanxi, China. *Catena* 39, 69-78.
- 332 Fu, B., Hu, C., Chen, L., Honnay, O., Gulinck, H., 2006. Evaluating change in agricultural landscape
333 pattern between 1980 and 2000 in the Loess hilly region of Ansai County, China. *Agriculture,
334 Ecosystems & Environment* 114, 387-396.
- 335 Gholami, L., Banasik, K., Sadeghi, S.H., Khaledi Darvishan, A., Hejduk, L., 2014. Effectiveness of
336 straw mulch on infiltration, splash erosion, runoff and sediment in laboratory conditions. *Journal of
337 Water and Land Development* 22.
- 338 Gholami, L., Sadeghi, S.H., Homae, M., 2013. Straw mulching effect on splash erosion, runoff, and
339 sediment yield from eroded plots. *Soil Science Society of America Journal* 77, 268-278.
- 340 Gómez, J.A., Romero, P., Giráldez, J.V., Fereres, E., 2004. Experimental assessment of runoff and soil
341 erosion in an olive grove on a Vertic soil in southern Spain as affected by soil management. *Soil Use
342 and Management* 20, 426-431.
- 343 Hammel, J.E., Papendick, R.I., Campbell, G.S., 1981. Fallow tillage effects on evaporation and
344 seedzone water content in a dry summer climate. *Soil Science Society of America Journal* 45, 1010-
345 1022.



- 346 Holland, J.M., 2004. The environmental consequences of adopting conservation tillage in Europe:
347 reviewing the evidence. *Agriculture, Ecosystems & Environment* 103, 1-25.
- 348 Huang, J., Wu, P., Zhao, X., 2013. Effects of rainfall intensity, underlying surface and slope gradient on
349 soil infiltration under simulated rainfall experiments. *Catena* 104, 93-102.
- 350 Hueso-González, P., Ruiz-Sinoga, J.D., Martínez-Murillo, J.F., Lavee, H., 2015. Overland flow
351 generation mechanisms affected by topsoil treatment: Application to soil conservation. *Geomorphology*
352 228, 796-804.
- 353 IBM, C., 2010. IBM SPSS Statistics for Windows, Version 19.0, IBM Corporation, Armonk, NY.
- 354 Jordán, A., Zavala, L.M., Gil, J., 2010. Effects of mulching on soil physical properties and runoff under
355 semi-arid conditions in southern Spain. *Catena* 81, 77-85.
- 356 Kang, S., Zhang, L., Song, X., Zhang, S., Liu, X., Liang, Y., Zheng, S., 2001. Runoff and sediment loss
357 responses to rainfall and land use in two agricultural catchments on the Loess Plateau of China.
358 *Hydrological Processes* 15, 977-988.
- 359 Katuwal, S., Vermang, J., Cornelis, W.M., Gabriels, D., Moldrup, P., de Jonge, L.W., 2013. Effect of
360 Root Density on Erosion and Erodibility of a Loamy Soil Under Simulated Rain. *Soil Science* 178, 29-
361 36.
- 362 Kukal, S., Sarkar, M., 2010. Splash erosion and infiltration in relation to mulching and polyvinyl alcohol
363 application in semi-arid tropic. *Archives of Agronomy and Soil Science* 56, 697-705.
- 364 Lal, R., 1993. Tillage effects on soil degradation, soil resilience, soil quality, and sustainability. *Soil and*
365 *Tillage Research* 27, 1-8.



- 366 Lee, H., Yang, Y., 1965. Effect of summer plowing on the winter wheat. *Acta Pedologica Sinica* 13,
367 405-411 (in Chinese with English abstract).
- 368 Lipiec, J., Kuś, J., Słowińska-Jurkiewicz, A., Nosalewicz, A., 2006. Soil porosity and water infiltration
369 as influenced by tillage methods. *Soil and Tillage Research* 89, 210-220.
- 370 Llorens, P., Domingo, F., 2007. Rainfall partitioning by vegetation under Mediterranean conditions. A
371 review of studies in Europe. *Journal of Hydrology* 335, 37-54.
- 372 Moreno-Ramón, H., Quizembe, S.J., Ibáñez-Asensio, S., 2014. Coffee husk mulch on soil erosion and
373 runoff: experiences under rainfall simulation experiment. *Solid Earth* 5, 851-862.
- 374 Mulumba, L., Lal, R., 2008. Mulching effects on selected soil physical properties. *Soil & Tillage
375 Research* 98, 106-111.
- 376 NDRC (National Development and Reform Commission), MWR (Ministry of Water Resources of
377 China), MOA (Ministry of Agriculture of China), SFA (State Forestry Administration of China), 2010.
378 Outline of integrated management on the Loess Plateau from 2010 to 2030.
- 379 Prosdocimi, M., Jordán, A., Tarolli, P., Keesstra, S., Novara, A., Cerdà, A., 2016. The immediate
380 effectiveness of barley straw mulch in reducing soil erodibility and surface runoff generation in
381 Mediterranean Vineyards. *Science of the Total Environment* 547, 323-330.
- 382 Puustinen, M., Koskiahho, J., Peltonen, K., 2005. Influence of cultivation methods on suspended solids
383 and phosphorus concentrations in surface runoff on clayey sloped fields in boreal climate. *Agriculture
384 Ecosystems and Environmet* 105, 565-579.



- 385 Roth, C., Helming, K., 1992. Dynamics of surface sealing, runoff formation and interrill soil loss as
386 related to rainfall intensity, microrelief and slope. *Journal of Plant Nutrition and Soil Science* 155, 209-
387 216.
- 388 Sadeghi, S.H.R., Gholami, L., Sharifi, E., Khaledi Darvishan, A., Homaei, M., 2015. Scale effect on
389 runoff and soil loss control using rice straw mulch under laboratory conditions. *Solid Earth* 6, 1-8.
- 390 Shen, H., Zheng, F., Wen, L., Han, Y., Hu, W., 2016. Impacts of rainfall intensity and slope gradient on
391 rill erosion processes at loessial hillslope. *Soil and Tillage Research* 155, 429-436.
- 392 Shi, H., Shao, M., 2000. Soil and water loss from the Loess Plateau in China. *Journal of Arid*
393 *Environments* 45, 9-20.
- 394 Shinohara, Y., Otani, S., Kubota, T., Otsuki, K., Nanko, K., 2016. Effects of plant roots on the soil
395 erosion rate under simulated rainfall with high kinetic energy. *Hydrological Sciences Journal* 61, 2435-
396 2442.
- 397 Song, R., Wu, C., Mou, J., Jiang, Y., Guo, J., 2002. Effects of maize stubble remaining in field on
398 dynamics of soil microbial biomass C and soil enzyme activities. *Chinese Journal of Applied Ecology*
399 13, 303-306. (in Chinese with English abstract).
- 400 Strudley, M., Green, T., Ascoughii, J., 2008. Tillage effects on soil hydraulic properties in space and
401 time: State of the science. *Soil and Tillage Research* 99, 4-48.
- 402 Swella, G.B., Ward, P.R., Siddique, K.H.M., Flower, K.C., 2015. Combinations of tall standing and
403 horizontal residue affect soil water dynamics in rainfed conservation agriculture systems. *Soil and*
404 *Tillage Research* 147, 30-38.
- 405 Systat, 2008. *Sigma Plot*. Version 10.0, Systat Software Inc., San Jose, CA.



- 406 Tang, K., 2004. Soil and water conservation in China. Chinese Science Press: Beijing, p15-20 (in
407 Chinese).
- 408 Todd, R.W., Klocke, N.L., Hergert, G.W., Parkhurst, A.M., 1991. Evaporation from soil influenced by
409 crop shading, crop residue, and wetting regime. Trans. of ASAE 34, 461-466.
- 410 Van, K.O., Quine, T.A., Govers, G., Gryze, S.D., Six, J., Harden, J.W., Ritchie, J.C., McCarty, G.W.,
411 Heckrath, G., Kosmas, C., Giraldez, J.V., Silva, J.R.M.d., Merckx, R., 2007. The impact of agricultural
412 soil erosion on the global carbon cycle. Science 318, 626-629.
- 413 Vermang, J., Norton, L.D., Huang, C., Cornelis, W.M., da Silva, A.M., Gabriels, D., 2015.
414 Characterization of soil surface roughness effects on runoff and soil erosion rates under simulated
415 rainfall. Soil Science Society of America Journal 79, 903-916.
- 416 Wang, L., Dalabay, N., Lu, P., Wu, F., 2016. Effects of tillage practices and slope on runoff and erosion
417 of soil from the Loess Plateau, China, subjected to simulated rainfall. Soil & Tillage Research
418 <http://dx.doi.org/10.1016/j.still.2016.09.007>.
- 419 Won, C.H., Choi, Y.H., Shin, M.H., Lim, K.J., Choi, J.D., 2012. Effects of rice straw mats on runoff and
420 sediment discharge in a laboratory rainfall simulation. Geoderma 189-190, 164-169.
- 421 Wu, Q., Wang, L., Wu, F., 2014. Tillage - impact on infiltration of the Loess Plateau of China. Acta
422 Agriculturae Scandinavica Section B-Soil and Plant Science 64 (4). 341-349.
- 423 Yair, A., Lavee, H., 1976. Runoff generative process and runoff yield from arid talus mantled slopes.
424 Earth Surface Processes 1,235-247.



425 Zhao, L., Wang, L., Liang, X., Wang, J., Wu, F., 2013. Soil surface roughness effects on infiltration
426 process of a cultivated slopes on the Loess Plateau of China. *Water Resources Management* 27, 4759-
427 4771.

428 Zhao, X., Huang, J., Wu, P., Gao, X., 2014. The dynamic effects of pastures and crop on runoff and
429 sediments reduction at loess slopes under simulated rainfall conditions. *Catena* 119, 1-7.

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449 **Tables**

450 Table 1 Selected chemical and physical properties of the soil

Soil Type	Particle Size %			Soil Texture	Organic Matter %	pH	Wet Aggregate stability mm	CEC (Cation Exchange Capacity) cmol kg ⁻¹	CaCO ₃ g kg ⁻¹
	Sand	Silt	Clay						
Lou soil	30.0	43.7	26.3	clay loam	1.33	8.2	1.4	18.1	74.6

451

452 Table 2 The time to runoff initiation, runoff rate, runoff volume, sediment concentration and loss, and

453 cumulative infiltration amount for WS and TP under each condition

Slope	treatment	Time to runoff (min)	Runoff rate (mm min ⁻¹)	Runoff volume (mm)	Sediment concentration (g L ⁻¹)	Sediment loss (g m ⁻²)	Cumulative infiltration amount(mm)
5	WS	23.68±4.29a	0.09±0.00b	3.00±0.51b	0.82±0.19b	2.41±0.14b	76.12±0.74a
	TP	4.78±0.36b	0.66±0.01a	37.86±0.67a	8.18±0.04a	304.31±5.93a	43.90±1.11b
10	WS	11.26±2.35a	0.07±0.00b	3.39±0.15b	0.89±0.17b	3.04±0.49b	75.95±0.54a
	TP	3.21±0.44b	0.83±0.03a	47.28±1.46a	10.29±0.91a	484.15±5.52a	32.53±2.01b
15	WS	7.07±2.78a	0.08±0.00b	4.16±0.22b	1.01±0.37b	3.78±0.76b	75.83±2.86a
	TP	2.99±0.09a	0.90±0.01a	51.42±0.69a	14.90±1.26a	731.23±11.66a	28.14±0.59b

454 For each slope gradient, the value is expressed as the mean ± the standard deviation, where the same lowercase

455 letters indicate no significant difference.

456 **Figures**

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458



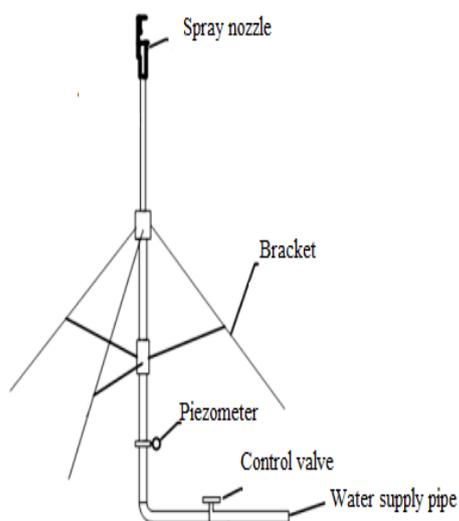
459 Figure 1 Experimental runoff pot (a) and schematic of the rainfall simulator (b).



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(a)



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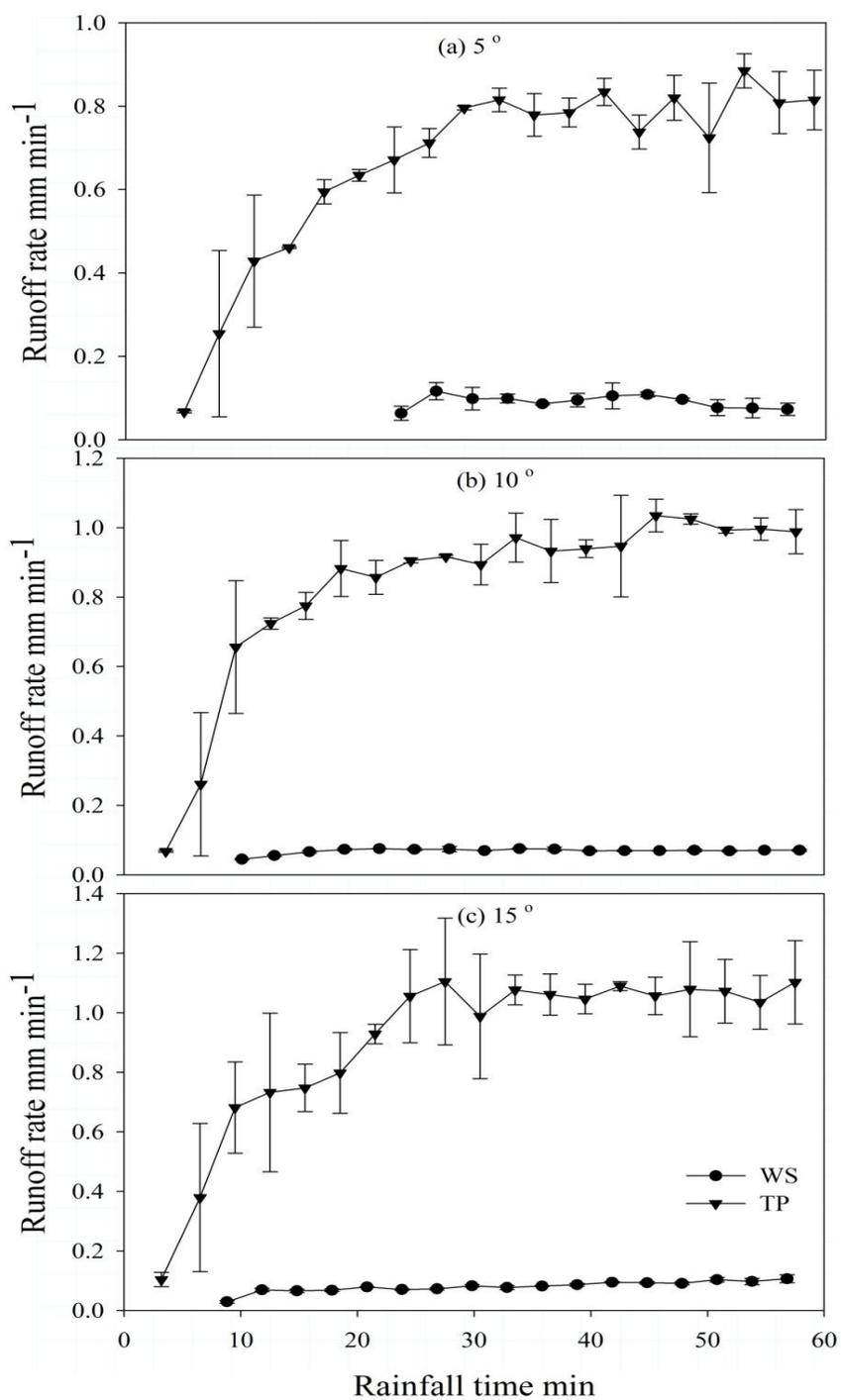
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(b)



465 Figure 2 Dynamics of the runoff rate in WS and TP plot during the simulated rainfall event .

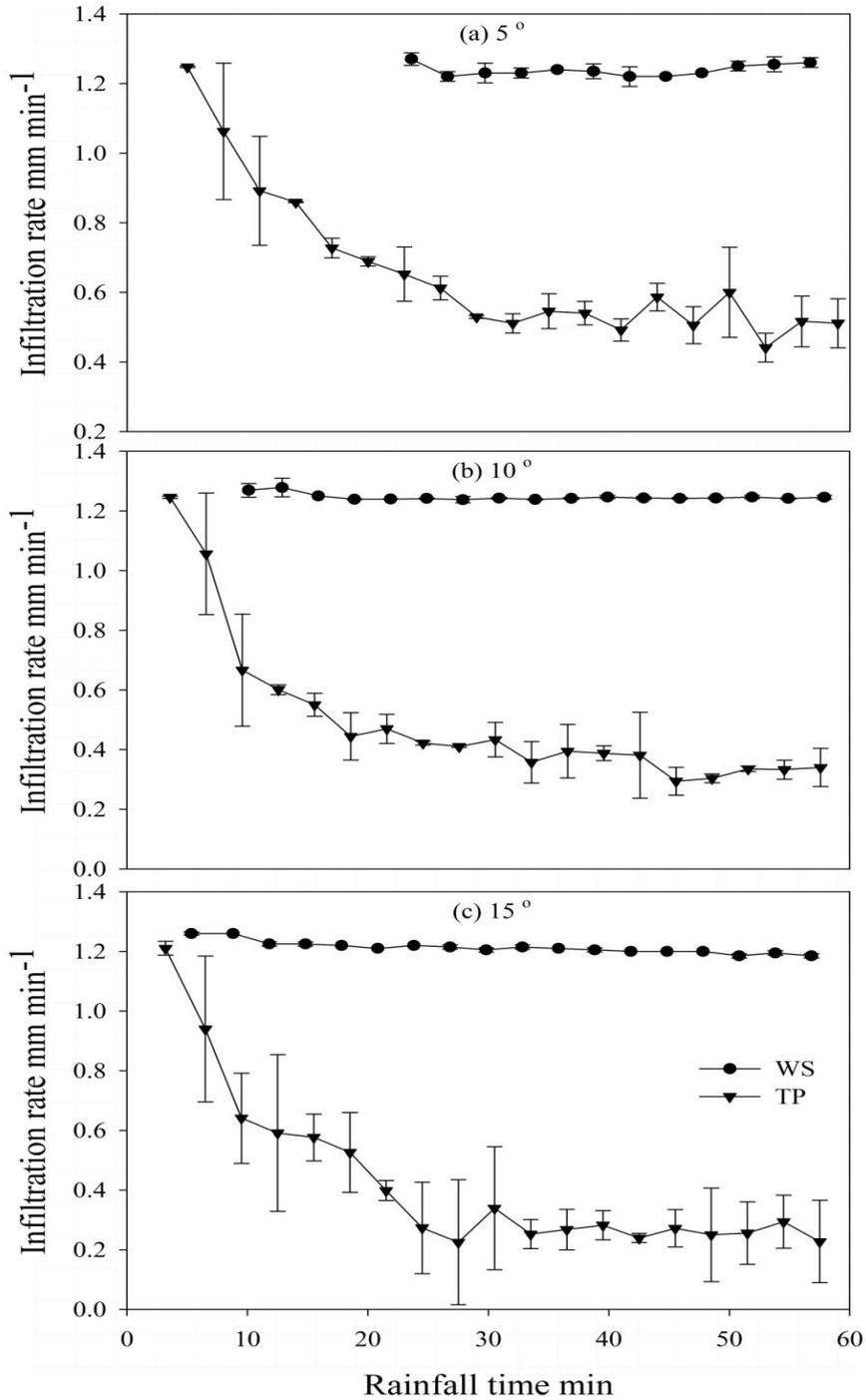


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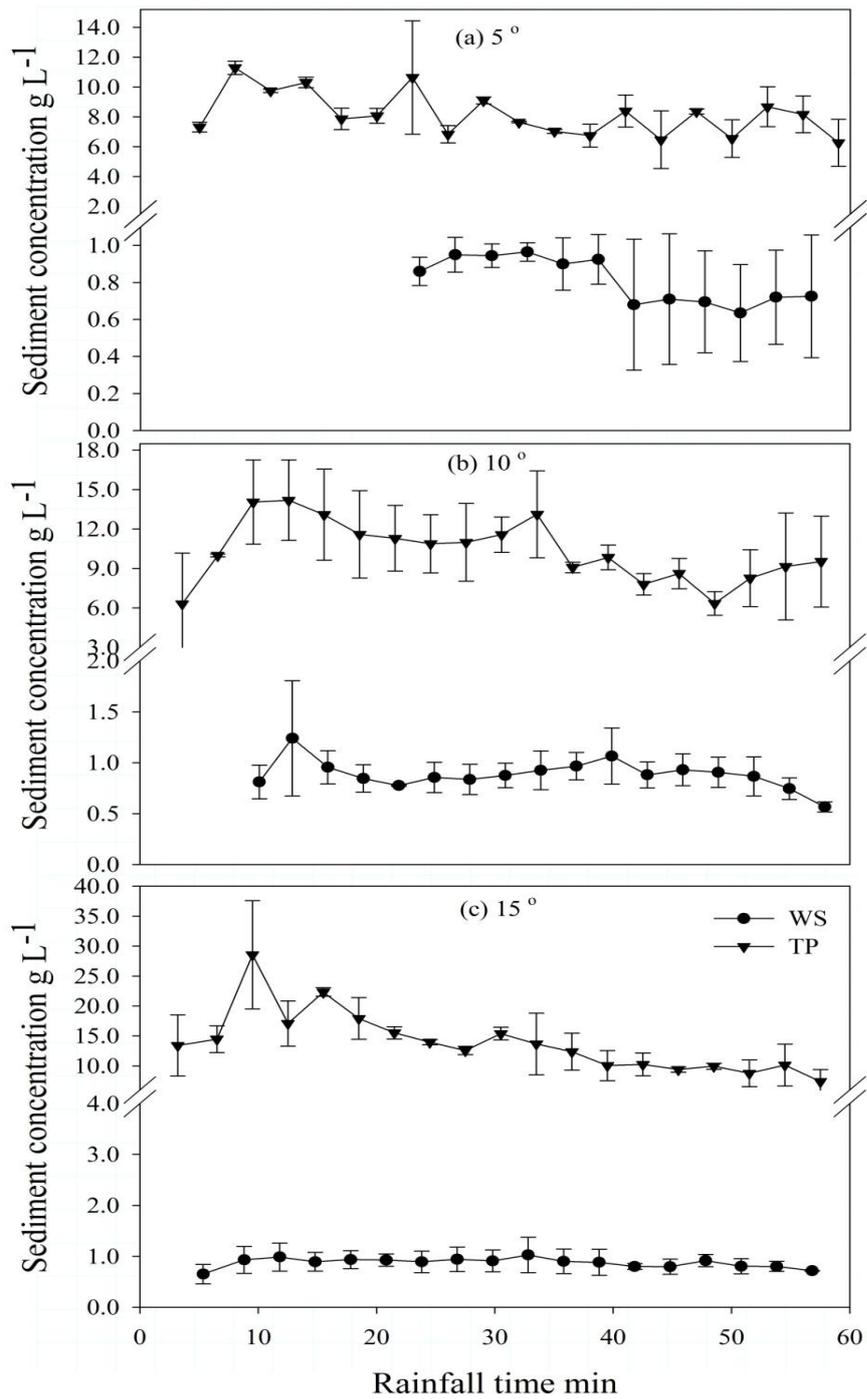
468 Figure 3 Dynamics of the infiltration rate in WS and TP plot during the simulated rainfall event.



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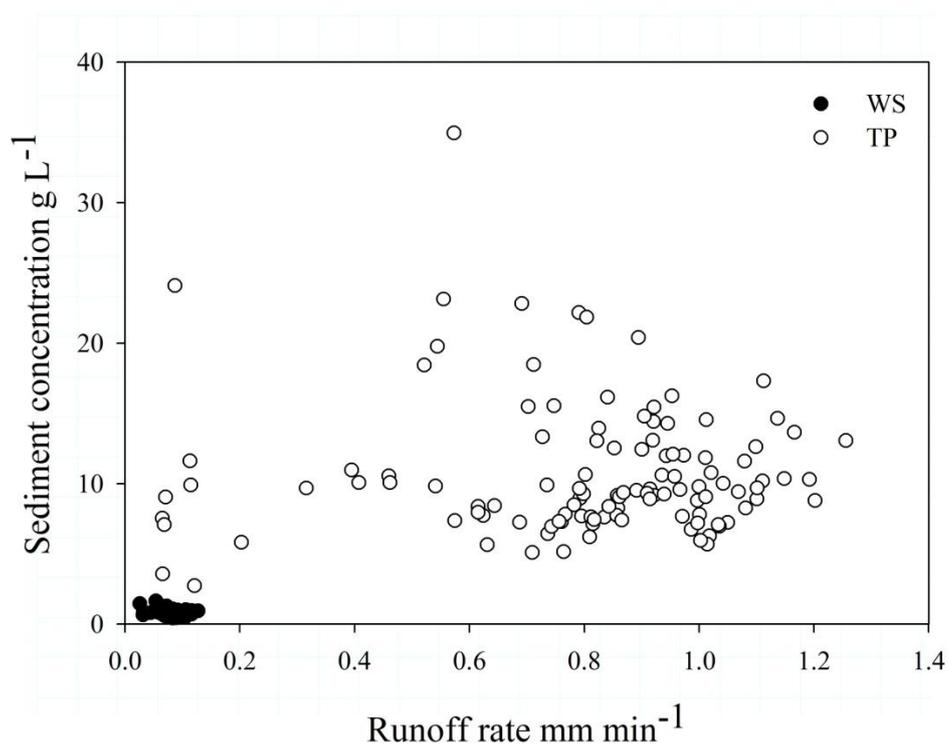
470 Figure 4 Dynamics of the Sc in WS and TP plot during the simulated rainfall event.



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472 Figure 5 Runoff rates versus sediment concentrations for WS and TP management treatment.



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